

# **Non-invasive measurement of steroids in fish-holding water: important considerations when applying the procedure to behaviour studies**

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## **Summary**

Fish behaviourists are increasingly turning to non-invasive measurement of steroid hormones in holding water, as opposed to blood plasma. When some of us met at a workshop in Faro, Portugal, in September, 2007, we realised that there were still many issues concerning the application of this procedure that needed resolution, including: Why do we measure release rates rather than just concentrations of steroids in the water? How does one interpret steroid release rates when dealing with fish of different sizes? What are the merits of measuring conjugated as well as free steroids in water? In the 'static' sampling procedure, where fish are placed in a separate container for a short period of time, does this affect steroid release — and, if so, how can it be minimised? After exposing a fish to a behavioural stimulus, when is the optimal time to sample? What is the minimum amount of validation when applying the procedure to a new species? The purpose of this review is to attempt to answer these questions

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and, in doing so, to emphasize that application of the non-invasive procedure requires more planning and validation than conventional plasma sampling. However, we consider that the rewards justify the extra effort.

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## Introduction

The non-invasive procedure for measuring steroids in fish-holding water has recently been comprehensively reviewed (Scott & Ellis, 2007). Briefly, the concentrations of steroids (and their metabolites) in the blood are normally used to study the endocrinological status of individuals and how they respond to a variety of physiological, behavioural and environmental factors. However, blood sampling of fish may be problematic when fish are too small, too rare or too valuable to be bled. Also, in order to be bled, they have to be netted, exposed to air, anaesthetised and handled — all of which can alter the physiological status and behaviour not only of the individual that is being sampled, but also of other fish in the tank. One way to overcome some of these problems is to ‘non-invasively’ measure the amounts of steroid that are released by fish into water. In terrestrial vertebrates, this has primarily been achieved by measurement of the rate of excretion of free (i.e., without any conjugating group such as a sulphate or glucuronide) or conjugated steroid metabolites in urine (Graham, 2004), faeces (Palme, 2005), saliva (Beerda *et al.*, 1996) or hair (Koren *et al.*, 2002). Although measurement of steroids has also been applied to fish urine (Oliveira *et al.*, 2001) and faeces (Oliveira *et al.*, 1999; Turner *et al.*, 2003), a much simpler and more direct approach is available. This is based on the fact that, in fish, free steroids are able to diffuse from the bloodstream into the water through the gills (Cravedi *et al.*, 1993; Vermeirssen & Scott, 1996; Sorensen *et al.*, 2000; Ellis *et al.*, 2005). Because this process appears to happen just as easily in reverse (Vermeirssen & Scott, 1996; Scott *et al.*, 2005; Maunder *et al.*, 2007), one hesitates to refer to it as an excretory process in the same way as urination or defecation. A more apt term is probably ‘leakage’. However, we will refer to it as ‘release’, as it is not clear yet whether the passage of all free steroids across the gills is incidental or active — conferring some advantage to the fish.

Since free steroids in the water seem mainly to be derived by passive diffusion across the gills, their rate of release should, in theory, be directly proportional to their concentration in plasma (i.e., steroid amount in water

should be a direct proxy for steroid concentrations in plasma). So far, the theory has been borne out in practice. In at least nine studies, a positive relationship was found between the amounts of free steroid released into the water over a given period of time and the concentration of free steroid in the plasma: 17,20 $\beta$ -dihydroxypregnen-4-en-3-one (17,20 $\beta$ -P) in goldfish (*Carassius auratus*; Stacey et al., 1989); 17,20 $\beta$ -P in dentex (*Dentex dentex*; Greenwood et al., 2001); cortisol in rainbow trout (*Oncorhynchus mykiss*; Ellis et al., 2004; Scott & Ellis, 2007); cortisol in Atlantic salmon (*Salmo salar*; Ellis et al., 2007b); 11-ketotestosterone (11-KT), androstenedione (Ad) and cortisol in stickleback (*Gasterosteus aculeatus*; Sebire et al., 2007); and cortisol in the convict cichlid (*Cryptoheros nigrofasciatus*; Wong et al., 2008: this issue). It has also been shown that the patterns of release of 17,20 $\beta$ -P and testosterone (T) by female goldfish (Scott et al., 2001) and of cortisol by sea bass (*Dicentrarchus labrax*; Fanouraki et al., 2008: this issue) closely matched changes in plasma steroid concentrations recorded in separate experiments. There has been only one published study in which a statistically significant relationship was not found — with 11-KT in the Siamese fighting fish (*Betta splendens*); however, sample numbers in this study were low (Dzieweczynski et al., 2006).

### **How does one assess and interpret steroid release?**

At first glance, the non-invasive procedure in fish may seem very simple: instead of measuring steroid concentration in a single blood sample, one measures it in a sample from the tank or pond that the fish inhabits. However, this procedure rarely yields usable data because, with only a single point measurement of steroid concentration, it is not possible to disentangle how much steroid has been released in the present, how much in the past, and how much has been lost due to water replacement, degradation or re-absorption by the fish (i.e., one cannot infer the hormonal status of the fish from this single point measurement). The critical information that is needed to make judgements on the hormonal status of the fish is how much steroid the fish release over a unit period of time (referred to in this paper as the 'steroid release rate'). At the moment, it is only possible to do this in two ways (see references in Scott & Ellis, 2007): (1) the 'static sampling procedure' which involves temporarily removing a fish from its tank, placing it in

clean water in a sampling container for a fixed period of time; (2) the ‘dynamic sampling procedure’ which involves keeping the fish in carefully controlled flow-through water conditions, and taking at least two water samples in order to estimate steroid release over the time between the samples. Both methods integrate steroid release rate over time and, thus, tend to ‘smooth out’ fluctuations that may occur in plasma concentrations — a feature that is considered an advantage of the non-invasive procedure (Goymann, 2005). The static sampling procedure has typically been used in behavioural studies as it is simpler to perform and enables measurement of individual steroid release rates where two or more fish are present in the same tank (i.e., in a social context), but the handling presents potential problems that are discussed later in this paper.

A common misconception about the non-invasive procedure is that steroid release rates have an absolute correspondence to plasma steroid concentrations (i.e., that a release rate of  $x$  ng/g per h will always correspond to a plasma concentration of  $y$  ng/ml whatever the steroid, the study or the species). However, this is far from the truth. Although steroid release rates are strongly correlated to plasma steroid concentrations (as described above), the intercepts (and sometimes probably the slopes) of the relationships are dependent on a wide variety of factors:

(1) Gill surface area: mass-specific gill surface areas in fish can vary 100-fold between species (Graham, 2006). On the basis that free steroids mainly diffuse across the gills, one could expect similar differences in steroid release rates across species or even within species across size classes. The potential effect of size is discussed in its own section below. Oxygen availability has also been shown to influence gill surface area in cyprinids (Nilsson, 2007).

(2) Changes in gill permeability and integrity: stress has been suggested to alter gill permeability. In rainbow trout, for example, it has been shown (Sloman *et al.*, 2004) that subordinate (and presumably stressed) individuals have a higher throughput of water and increased sodium efflux across the gills compared to dominant individuals. At the moment it is not known whether this would also influence steroid release rate, but this possibility should be borne in mind when applying the non-invasive procedure to studies on fish social behaviour. It is also likely that the rate of free steroid release may be altered by water quality factors and diseases that affect the integrity of the gills (Ellis *et al.*, 2004, 2007a) although this has yet to be proved experimentally.

(3) Binding affinity of steroids for sex hormone binding globulin (SHBG) in plasma: steroids for which SHBG has a high affinity and capacity (e.g., T and E<sub>2</sub>) should in theory be released at a slower rate than those steroids that do not (e.g., 11-KT and cortisol). The basis for this theory is that, when a steroid is bound to SHBG, it is effectively removed from solution and, thus, lowers the diffusion gradient across the gills. There is a certain amount of circumstantial evidence to support this assumption. Comparison of a range of species indicates that values for E<sub>2</sub> and T tend to fall at the low end of the range of steroid release rates (Scott & Ellis, 2007). Also, though the concentration of T is higher than that of 17,20 $\beta$ -P in the plasma of female goldfish, the opposite is true for the amounts that are released into the water (Scott et al., 2001; Scott & Ellis, 2007). Furthermore, the rates of uptake of several steroids from the water by tench (*Tinca tinca*) were found to correlate positively with their ability to bind to SHBG (Scott et al., 2005). Since the amount of SHBG can vary markedly between species (Hobby et al., 2000b) and stage of sexual maturation (Hobby et al., 2000a) one might predict a marked effect on the rate of release of strongly-bound steroids such as E<sub>2</sub> and T, but little or no effect for poorly-bound steroids such as cortisol and 11-KT. This hypothesis has yet to be confirmed by experimental investigation, although preliminary work has shown that it is possible to predict how much steroid is present in an unbound (i.e., diffusible) form from information on the affinity and binding capacity of SHBG and the total amount of free steroid in the plasma (Scott et al., 2005).

(4) Steroid lipophilicity: when dogfish (*Squalus acanthias*) were injected with eight drugs with widely differing chemical properties (Maren et al., 1968), the lipid-insoluble ionised drugs (cf., sulphated and glucuronidated steroids) were excreted via the kidney, while the lipid-soluble non-polar drugs (cf., free steroids) diffused freely into the water across the gills; similarly, when free 17,20 $\beta$ -P, sulphated 17,20 $\beta$ -P and glucuronidated T were injected into rainbow trout, only the free 17,20 $\beta$ -P was cleared via the gills (Vermeirssen & Scott, 1996). What is not yet clear, however, is whether more subtle differences in steroid polarity are sufficient to cause differences in their relative rates of release via the gills. For example, for two steroids of similar polarity (17-hydroxyprogesterone and Ad) the rates of uptake were shown to be widely different, although consistent with differences in how strongly they bound to SHBG (Scott et al., 2005).

(5) Conversion of free steroids: it is likely that most free steroid found in the water will have followed a very simple pathway: synthesis in the gonads (sex steroids) or interrenal tissue (cortisol) → release into the bloodstream → release into the water by diffusion across the gills. However, while in the bloodstream, it is very likely that some free steroids will undergo conversion to other free steroids. It is known, for example, that fish blood cells contain a  $17\beta$ -hydroxysteroid dehydrogenase ( $17\beta$ -HSD) that readily converts 11-ketoandrostenedione made in the testes to 11-KT (Mayer et al., 1990) and the liver of the African catfish (*Clarias gariepinus*) converts testicular  $11\beta$ -hydroxytestosterone into 11-KT (Cavaco et al., 1997). Also, cortisol is rapidly converted peripherally to cortisone in fish (Pottinger & Moran, 1993). The importance of such conversions in regulating the amounts of free steroids that can be measured in water (and plasma) is something that needs to be examined in future studies.

(6) Restriction of movement: rainbow trout held in a box with a gill dam released less [ $^{14}\text{C}$ ]methyl testosterone into the water (38% of the injected amount) than 'free-swimming' fish held in a 5 l aquarium (53%) (Cravedi et al., 1989). Restricting the movement of dogfish also reduced the branchial release of lipid-soluble drugs (Maren et al., 1968). However, the release rates of free cortisol and melatonin by rainbow trout held in a box with a gill dam were similar to those calculated for free-swimming fish (Ellis et al., 2005).

(7) Involvement of other routes of release: although the gills appear to be the major source of free steroid release in fish (at least based on the four previously-mentioned studies carried out on two species — rainbow trout and goldfish), they are not the only source. Urine also appears to contain small amounts of free steroids (though in most cases <10% of conjugated steroid): cortisol and  $17,20\beta$ -P in Pacific herring (*Clupea harengus*) and plaice (*Pleuronectes platessa*; Scott et al., 1991; Scott & Canario, 1992), T (Moore & Scott, 1991) and  $17,20\beta$ -P (Moore & Scott, 1992) in Atlantic salmon; T and  $17,20\beta$ -P in rainbow trout (Scott & Liley, 1994; Vermeirssen & Scott, 1996); Ad,  $17,20\beta$ -P and  $17,20\beta,21$ -trihydroxypregn-4-en-3-one ( $17,20\beta,21$ -P) in goldfish (Sorensen et al., 2000); and 11-KT in male tilapia (*Sarotherodon galilaeus*; Hirschenhauser et al., 2008: this issue). In this final study, concentrations of free 11-KT in urine were significantly correlated to those in plasma. Alternatively, free steroids may find their way into the water through the bile (Pottinger et al., 1992), mucus (Schultz et al., 2005), milt (Scott et al., 1991), ovarian fluid (Schoonen et al., 1989), seminal vesicle

fluid (Schoonen et al., 1988) and possibly skin (Ellis et al., 2005). However, no one has yet attempted to gauge the relative importance of these other pathways compared to the gills. If such routes do contribute significant amounts of free steroid to water samples, then active events such as urination and/or defecation might compromise results.

The above list is not comprehensive. Other factors that potentially alter the rate of release of free steroids (or at least the amounts that can be measured in the water) include temperature, the propensity of the fish to reabsorb the steroids, metabolism and adsorption of the steroids (by biota and container walls).

The existence of factors other than plasma steroid concentration that are able to regulate the rate of release of free steroids should not deter anyone from applying the non-invasive procedure. All it means is that careful thought needs to be given to experimental design so that the factors that affect release rate are controlled (or, at the very least, acknowledged and considered in the interpretation of results).

### **How does one take fish size into account in experiments where fish have markedly different sizes?**

It is standard practice to correct physiological measurements related to the metabolism of the focal animal by adjusting for the body size of the individual (West & Brown, 2005). Therefore, in most published studies, steroids released into the water (per unit of time) have been scaled to body mass — i.e., dividing amount of steroid released into the water by body mass (Scott & Ellis, 2007). This is considered a valid approach when all fish are of a roughly similar size. However, where fish are of markedly different sizes, the potential effect of body size on the release process needs to be considered.

For any 3-dimensional object with fixed relative dimensions, i.e., a constant shape, an increase in length by a factor of  $x$  will be accompanied by an  $x^2$  increase in surface area and an  $x^3$  increase in volume ( $\approx$ mass). Thus, the surface area of the gills would be expected to be proportional to  $\text{mass}^{2/3}$ . To illustrate this, when a fish grows from 4 to 10 g (i.e., an increase of 250%), the gill surface area increases by only 184%. On this basis, a 10-g fish should release steroid at roughly twice the rate as a 4-g fish. There are only two

studies (Bender et al., 2006; Bender et al., 2008), both on *Neolamprologus pulcher*, that have applied the non-invasive procedure to fish of markedly different sizes (between 3 and 12 g). The data in these studies provided no evidence that fish of ca. 10 g released twice as much steroid as fish of ca. 4 g — in fact, both size groups seemed to release very similar amounts of steroid over the same time interval. These results, however, should not be accepted as evidence that steroid release rate does not increase in relation to body size. This is because a key bit of information is missing from these studies, viz., plasma steroid concentrations. The question about whether steroid release rate is related to body size will only be answered when experiments have been carried out that include both body size and plasma steroid concentrations as covariates.

### **Is it worth measuring conjugated steroids?**

Over the last ten years, researchers applying the non-invasive procedure in fish to behavioural studies have diverged into two ‘schools’ — those measuring free steroids only and those measuring the free, sulphated and glucuronidated steroid fraction (and more often than not reporting the combined levels as ‘total’ steroid).

A disadvantage of measuring conjugated steroids is that their pathways from synthesis to release are far more complicated than in the case of free steroids. Sulphation and glucuronidation of steroids can take place not just in the liver (as a primary means of deactivating the steroids and making them more water soluble) but also directly in the gonads of some species (Scott & Vermeirssen, 1994). Additional complexity is conferred by the fact that the conjugated steroids can be temporarily ‘stored’ in the bile and urine and their release is, thus, subject to factors such as glomerular filtration rate, urination frequency, feeding, gut passage time and defecation.

Only a single study has been carried out on the specific routes of release of sulphated vs. glucuronidated steroids in teleost fish. That study showed that tritiated 17,20 $\beta$ -P-sulphate injected into the blood stream of rainbow trout was mainly excreted via the urine, whereas glucuronidated T was mainly transferred to the bile (Vermeirssen & Scott, 1996). However, one should be cautious in generalising this finding to other fish and to other steroids. In goldfish, for example, equal amounts of glucuronidated and sulphated



17,20 $\beta$ -P could be found in the urine (Sorensen et al., 2000), whereas when rainbow trout were injected with tritiated free cortisol (Idler & Truscott, 1972), little or no radioactivity was recovered in the urine.

A perceived advantage of measuring all three moieties (i.e., free, sulphated and glucuronidated) is that one is measuring 'total' steroid release. However, this does not consider the fact that steroids are often metabolised into other steroids in addition to being conjugated. In rainbow trout, for example, little or no cortisol was found in the conjugated steroid fraction of the bile, but there were very large amounts of 5 $\beta$ -reduced metabolites of both cortisol and cortisone (Pottinger et al., 1992). Also, in reproductively mature male and female plaice (Scott et al., 1998; Vermeirssen et al., 1998) and wolffish (*Anarhichas lupus*; Tveiten et al., 2000) the amounts of sulphated 5 $\beta$ -reduced 17,20 $\beta$ -P in plasma and/or urine hugely exceeded the amounts of intact 17,20 $\beta$ -P in all three fractions combined.

Despite the complexities of origin, metabolism and routes of excretion of conjugated steroids, in the few studies where they have been measured in water, their rates of excretion (either measured directly in urine or in the water) have yielded remarkably consistent and meaningful data. For example, conjugated T, 17,20 $\beta$ -P and 17,20 $\beta$ ,21-P in gonadotropin-injected female goldfish (Scott & Sorensen, 1994) and T in male goldfish (Sorensen et al., 2005) were released in a regular pattern that in some cases followed the free steroid peaks by a few hours; the levels of conjugated T and 17,20 $\beta$ -P in female rainbow trout urine followed the pattern of changes of T and 17,20 $\beta$ -P in plasma and stage of maturity of the fish (Scott & Liley, 1994); the rates of release of both sulphated and glucuronidated 17,20 $\beta$ -P by GnRH-injected dentex were positively correlated to plasma 17,20 $\beta$ -P concentrations (Greenwood et al., 2001) — though their coefficients of correlation were lower than for free 17,20 $\beta$ -P.

Situations when it might be considered an advantage to measure conjugated in addition to (or instead of) free steroids include:

- (1) Measurement of conjugated steroids has been shown to reveal behavioural and physiological strategies that would be missed (and would be difficult to continue to investigate) by measurement of free steroids alone. For example, the male peacock blenny (*Salaria pavo*) bears an accessory sex organ on its anal fin and exclusively releases sulphated 17,20 $\beta$ -P into the water (Oliveira et al., 1999). Also, males of some cichlid species urinate in response to social signals, e.g., when interacting with other males (Almeida

*et al.*, 2005; Barata *et al.*, 2007). These species — plus the types of behaviour that stimulates urine release — can be revealed by examining the ratio of sulphated to free steroid concentrations in water (Hirschenhauser *et al.*, 2008: this issue).

(2) Based on the time-release studies on male and female goldfish (mentioned above), the production of some conjugated steroids lags that of the free steroids and, thus, probably represents an estimate of what was happening to the fish several hours prior to the sampling. This fact could be useful for: field studies; for experiments where critical hormone changes take place at inconvenient times (e.g., at night); and for overcoming the ‘handling effect’ of the static sampling procedure (see below).

(3) Some conjugated steroids are used by fish as pheromones and, thus, they need to be measured to study how and when they are released, and to understand the basis for the behavioural/physiological reactions they elicit in ‘receivers’. In fact, this was the original impetus for measuring conjugated steroids in fish water (Stacey *et al.*, 1989; Scott & Sorensen, 1994; Sorensen & Scott, 1994).

In conclusion, in studies where it is the intention to use the non-invasive procedure as a direct replacement for blood sampling, we consider it sufficient to measure the free steroid fraction in water only. However, there will always be certain experiments and certain lines of research where it would be an advantage to also measure the conjugated steroids. In such situations, we recommend, however, that the different fractions are reported separately — not as ‘combined’ or ‘total’.

### **Does the static sampling procedure induce a ‘handling effect’ on steroid release?**

As stated above, the static sampling procedure (i.e., where the fish is taken out of its tank and placed in a small sampling container for 30 to 60 min) has been the method of choice for behavioural studies. It has to some extent been assumed that this mildly intrusive procedure would have little effect on steroid production and release by the fish. However, we now have to accept that the procedure constitutes a stressor, and that 30 min is more than enough time for some of the cortisol produced in response to this stress to be released into the container. For instance, there was a significant increase in cortisol

release (compared to controls) within 30 min in three different species that were subjected to ship noise while being held in a container (Wysocki et al., 2006). Experiments on salmon under flow-through conditions also showed a significant three-fold increase in cortisol concentrations within 30 min of handling stress (Ellis et al., 2007b). Dr Ryan Earley reported at the Faro Workshop (unpublished data) that if fish were transferred from their home tanks to sampling containers every day for 30 min, the amount of cortisol that they released decreased significantly over a 4-day period (i.e., they appeared to habituate to the handling and temporary confinement). This suggests one approach to diminishing, or possibly even abolishing, the handling effect on cortisol production. Another approach might be to make the length of the sampling period as short as possible (e.g., 10 min instead of 30 min). However, such a step could, by reducing the total amounts of steroid collected, limit the number of assays for other steroids that could be carried out on the extract. Yet a third approach (already mentioned above) is to measure a conjugate of cortisol or one of its metabolites, as these lag the production of free cortisol and, thus, take much longer than 30 min to respond to the handling stress (Pottinger et al., 1992).

Unfortunately, the static sampling procedure also appears to be able to influence the production and/or release of steroids other than cortisol. A perceived advantage of the non-invasive procedure is that a fish can be sampled both before and after a behavioural interaction and, thus, potentially act as its own control. However, it has been noted in a number of studies (Schreck, 1972; Pickering et al., 1987; Scott & Canario, 1990; Kubokawa et al., 1999; Vermeirssen et al., 2000) that fish sampling is followed within 30 to 60 min by a significant decrease in plasma androgen concentrations. A similar significant decrease was found in rate of release of 11-KT into water by male sticklebacks (Sebire et al., 2007) and in the concentration of 11-KT in the urine of male tilapia (Oliveira et al., 2005). In this latter study, a proposal was made that the drop might be due to a natural diurnal variation of androgen secretion. However, no diurnal pattern of either T or 11-KT release was found in the bluebanded goby (*Lythrypnus dalli*) (Lorenzi et al., 2008) and, in the two studies in which both possibilities were specifically tested (i.e., handling v. diurnal changes), it was only the repeatedly handled fish that showed a significant drop in plasma androgen concentrations (Schreck, 1972; Pickering et al., 1987). Another study on rainbow trout and brown trout

(*Salmo trutta*) suggested that the drop could be mimicked by implanting cortisol (Carragher et al., 1989). Whether or not stress associated with handling or confinement is the primary cause, procedures that involve 'before' and 'after' measurements should not be adopted until it has been clearly demonstrated whether or not handling reduces androgen (and indeed any steroid) production and, if so, how long it is before the fish recover their normal response. In the study in which brown trout were subjected to an acute stress, it took up to 24 h for T and 11-KT concentrations to return to their original concentrations (Pickering et al., 1987).

### **When should fish be sampled after exposure to a stimulus?**

When designing experiments in which one subjects a fish to a test stimulus (e.g., placing a male with a female) and then takes a sample (either blood or water) for steroid analysis, it is important to know how long it takes for the fish to mount a response. The pattern of response of most compounds to a stimulus follows a roughly bell-shaped curve, with the optimum time for sampling generally being when the curve reaches its peak. Unfortunately, this is something that can only be determined experimentally. In fish, the time to maximum steroid release rate (Table 1) varies depending on whether it is a reproductive steroid (between 2 and 18 h) or a stress steroid (between 0.5 and 4 h), whether it is the free or conjugated form of the steroid that is being measured, and whether it is an androgen or progestagen. The speed of response is probably also dependent on temperature and, in the case of the sex steroids, on the stage of maturation of the gonads, i.e., how much they are 'primed' to respond. It is also important to note that in a flow-through system, there will be a lag between the time that a steroid reaches its maximum release rate and the time that it reaches its maximum concentration in the water (Ellis et al., 2004; Lower et al., 2005).

### **What is the minimum amount of validation required for applying the non-invasive procedure to a new species?**

There are several different ways to firstly extract steroids from water samples and secondly to measure them. However, discussion of their relative merits is outside the scope of this present review. All the groups presently working

**Table 1.** Time that it takes for steroids to reach their maximum release rates after the fish have received a stimulus in water samples.

Species	Sex	Temperature (°C)	Stimulus	Stimulus duration	Steroid	Peak time post stimulus	Reference
Reproductive steroids							
Goldfish <i>Carassius auratus</i>	♀	18	Injection with gonadotropin	Once	17,20 $\beta$ -P Ad T	7.5 h (F), 9 h (G, S) 4 h (F) 4 h (F, G, S)	Scott & Sorensen (1994)
Tilapia <i>Oreochromis mossambicus</i>	♂	24	Watching two males fight	1 h	11-KT (in urine) T (in urine)	2 h (F+G+S) 30 min (F+G+S)	Oliveira et al. (2001)
Tench <i>Tinca tinca</i>	♂/♀	23	Injection with Ovaprim (GnRHa+ dopamine inhibitor)	Once	17,20 $\beta$ -P (♀)  11-KT (♂) Ad (♂)	18+ h (F) 24 h (G and S) 12 h 24 h	Piniillos et al. (2002)
Roach <i>Rutilus rutilus</i>	♀/♂	13–17	Injection with carp pituitary extract	Once	17,20 $\beta$ -P Ad	12–15 h (F, G) 3–6 h (F)	Lower et al. (2004)
Tilapia <i>Oreochromis mossambicus</i>	♂	28	Injection with GnRHa	Once	11-KT T	4 h (F) 2 h (F)	Hirschenhauser et al. (2004)
Goldfish <i>Carassius auratus</i>	♂	18	Injection with gonadotropin	Once	17,20 $\beta$ -P, 11-KT Ad, T T	4 h (F, G, S) 4 h (F) 8 h (G, S)	Sorensen et al. (2005)

Table 1. (Continued.)

Species	Sex.	Temperature (°C)	Stimulus	Stimulus duration	Steroid	Peak time post-stimulus	Reference
<b>Stress steroids</b>							
Rainbow trout <i>Oncorhynchus mykiss</i>	♀/♂	12	Single air exposure Repeat air exposure	90 s 90 s (at 0, 1 and 2 h)	Cortisol, Cortisone	0.5–1 h (F) 4–5 h (F)	Ellis et al. (2004) <sup>a</sup>
Carp <i>Cyprinus carpio</i>	♀/♂	17	Tag insertion	Once	Cortisol	1–2 h (F)	Lower et al. (2005) <sup>a</sup>
Roach <i>Rutilus rutilus</i>	♀/♂	10	Tag insertion	Once	Cortisol	2–4 h (F)	Lower et al. (2005) <sup>a</sup>
Atlantic salmon <i>Salmo salar</i>	♀/♂	15	Single air exposure	90 s	Cortisol	3 h (F)	Ellis et al. (2007b) <sup>a</sup>
European sea bass <i>Dicentrarchus labrax</i>	♀/♂	20	Chasing and air exposure Confinement (low density) Confinement (high density)	6 min 24 h 24 h	Cortisol	0.5–1 h (F) 1–2 h (F) 1 h (F)	Fanouraki et al. (2008): this issue <sup>a</sup>
Threadfin anthias <i>Pseudanthias squamipinnis</i>	♂	?	Injection with ACTH	Once (but fish handled every h)	Cortisol	4 h (F)	Bshary et al. (2007)

F, free (unconjugated); G, glucuronidated; S, sulphated.

<sup>a</sup> In flow-through systems, time of maximum steroid concentrations will lag behind time of maximum release rates.

in the field of fish behaviour use solid phase extraction of steroids followed by radio- or enzyme-immunoassay. Very few problems have been reported with the application of these procedures, and when they have (Ellis et al., 2004), they have been resolved. The most likely problem to be encountered is with the extraction from water of compounds that interfere with the immunoassays and yield false positive results (so-called 'matrix effects'). In some situations (e.g., when trying to measure steroids in large volumes or water samples collected in the field) this can be a major problem, the solution to which may be relatively difficult and expensive (Lopez de Alda & Barcelo, 2001). However, fish behaviour scientists generally have the luxury of being able to provide the fish with clean 'problem-free' water. Also, if using the static sampling procedure in conjunction with small fish, they only need to use small volumes of water (50–100 ml).

Whatever procedures are chosen, researchers should still validate their extraction procedure and assays. We recommend the following steps:

(1) Check during every experiment that the water provided to the fish has little or no background activity and/or interference.

(2) Add known and appropriate amounts of steroids to some water samples to check that there is an acceptable (80–100%) and consistent rate of recovery after extraction.

(3) Ensure there is parallelism between extracts and standards.

(4) Check the specificity of the procedure by separating water extracts on thin-layer or high-performance liquid chromatography (to verify that the steroid is recovered in its expected elution position).

(5) If water samples cannot be processed immediately, check for how long they can be left untouched, chilled or frozen without loss of activity.

(6) Check the 'robustness' of the procedure by giving the fish at least one physiological challenge to check that they respond in a biologically meaningful manner (i.e., a positive control). For example, injection of ACTH or application of acute stress should result in dose-dependent increases in the release of cortisol; GnRH or gonadotropin injection into sexually mature fish should also induce a dose-dependent increase in the release of sex steroids.

(7) If carrying out experiments involving physiological or behavioural challenges, carry out a time-course study to determine the optimum time between the experimental treatment and the peak of steroid release, noting that it is very likely that the peaks of androgens and progestagens (and free and conjugated steroids) will be at different times.

(8) If the experiment involves sampling the same fish more than once, it is also necessary to check whether there is any handling effect (or diurnal variation) that might affect the release of any of the steroids.

(9) If the experiment involves fish that fall into different size groups, it would be necessary to investigate release rates of steroids in relationship to steroid concentrations in plasma as well as to fish size.

Validation studies are very time consuming and expensive, and, thus, it is fair that people should ask the question when applying the procedure to a new species ‘does one have to carry out all the validation steps?’. Unfortunately, the correct answer is ‘most, if not all, of them’. There is huge variation in fish steroid biosynthetic pathways (Kime, 1993), steroid release rates (Scott & Ellis, 2007) and fish-holding conditions. With the field being relatively young, there is presently insufficient information to confidently state that a procedure that has been fully validated in one species — e.g., rainbow trout living in relatively clean and cool fresh water (Ellis *et al.*, 2004) — can be applied without problem to another that lives in organically rich and warm water. There may well be no problem. However, without proper validation, this can only be an assumption.

## Conclusion

The measurement of steroids in fish-holding water provides a non-invasive alternative to blood sampling, commensurate with the protocols for (and ethics of) behavioural experiments. However, the non-invasive approach requires a thorough appreciation of potential problems associated with sampling, processing, validation and interpretation of results. It requires, for example, the measurement of steroid release rates, and not just simple measurement of a steroid water concentration. Many factors (including probably fish size) can have a significant impact on steroid release rates and these must be controlled when designing experiments and interpreting the results. Although it may be useful to measure the sulphated and glucuronidated steroid fractions in water extracts, it should be sufficient in most cases to measure the free steroid only. The static sampling procedure has a drawback in that the capture and confinement in a sampling container constitutes a stressor to the fish. Even with a relatively short sampling period of 30 min, the fish mounts a detectable cortisol response, and the stress may also cause a temporary decrease in androgen production. Possible solutions to this problem



have been proposed. It is considered impossible to accurately predict the optimum time between application of a behavioural stimulus and maximum steroid release, and it, therefore, needs to be established experimentally at the start of any behavioural study. Although the extraction and assay of steroids does not appear to present a major methodological problem — especially when experiments are carried out in the laboratory where water quality is usually good — we still advise validation of the procedure (especially when applying it to new species and new water conditions).

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